ORIGIN OF SUBSOIL CARBON IN A CHINESE PADDY SOIL CHRONOSEQUENCE

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ABSTRACT. Rice paddies are highly important agricultural soils in view of their relevance as major staple food provider in the world and their key role in the global carbon cycle, caused by special management practices. A soil chronosequence, consisting of paddy and upland soils, developed on reclaimed estuarine sediments in the Province of Zhejiang, China, was sampled to investigate the influence of duration of agricultural use (50 to 2000 yr) on soil composition. The uniform composition of the parent material provides the unique opportunity to compare the effects of different land management practices (paddy and non-paddy) on soil carbon dynamics and the origin of organic carbon (OC) in top- and subsoils, using ¹⁴C measurements by accelerator mass spectrometry (AMS). The total soil organic carbon (TOC) was split into chemically defined pools of different mobility, namely the acid- and water-soluble fulvic acids (FA), the alkali-soluble humic acids (HA), and insoluble humin fraction. The more mobile HA and FA fractions contain significantly more ¹⁴C than the corresponding TOC and humin, indicating a downward transport of OC in the subsoil. Plant roots with ¹⁴C concentrations up to 128% of the modern standard, found far below the plough pan, reveal plant roots and root exudates as other direct sources of subsoil OC in paddy soils.

INTRODUCTION

Rice paddy soils are the most important agricultural soils in the world because (i) paddy soils provide the staple food for nearly half the world population (MacLean et al. 2002) and (ii) paddy soils play an important role in the global carbon and methane cycle. Rice is the major crop in the tropics and subtropics, whereas China represents the predominant rice producer with about 30% of the global rice production and 19% of the total world rice paddy area (FAOSTAT 2010). In the context of mitigating the increase of the atmospheric greenhouse gases CO_2 and CH_4 , paddy soils are quite important in view of the high soil organic carbon (SOC) level observed in paddy topsoils, their assumed high C sequestration potential (Pan et al. 2003; Xu et al. 2011), and the amenability of their properties to active management. The larger paddy topsoil SOC stocks, compared to upland soils, are caused by a large carbon input via straw incorporation and/or other organic additions (Tanji et al. 2003; Rui and Zhang 2010). Furthermore, the water-logged conditions during flooded phases lead to decreased decomposition rates and enhance the accumulation of SOC (Huang and Sun 2006; Kögel-Knabner et al. 2010).

After the oceans, soils are the largest active pool within the global carbon cycle (Lal 2008). The SOC pool is twice the size of the atmospheric C pool and nearly 3 times the size of the biotic C pool (Lal 2004). The amount of organic carbon (OC) in the upper 100 cm in the world's soils is estimated to be about 1550 Pg (1 Pg = 10^{15} g) (Batjes 1996). Since soils are the largest reservoir for C in the terrestrial ecosystem, they play a key role within the global carbon cycle, and SOC dynamics were examined in numerous studies of increasing organic complexity over the past decades (Scharpenseel et al. 1989; Trumbore et al. 1989; Becker-Heidmann and Scharpenseel 1992; Wang and Amundson 1996; Six et al. 2001; Baisden et al. 2002; Rethemeyer et al. 2005; Dreves et al. 2007; Flessa et al. 2008; Marschner et al. 2008; Laskar et al. 2012).

In many cases, only the topsoil (~0–30 cm) of the related soil profile is considered. Although those upper 30 cm contain <50% of the SOC (Jobbagy and Jackson 2000), the organic carbon located in subsoil horizons was assumed to be insignificant within the system of SOC transport and stabilization. Land-use-change related shifts in SOC stocks, for example, were assumed to be significant only in topsoils (O'Brien and Stout 1978; Baisden et al. 2002). The main reason for this point of

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view is the assumption that subsoil carbon is inert and does not take part in the dynamic interactions between different soil compartments. This is based on very low annual inputs of fresh organic matter and the limited spatial accessibility of subsoil OC for microbial activities and oxygen within the soil profile (Christensen 2001). Furthermore, in view of the low ¹⁴C activity of subsoil horizons, it is assumed that, with increasing depth, SOC becomes more stabilized and shows longer residence times (Rumpel and Kögel-Knabner 2011).

In the course of the discussion about global warming and the potential of deep-soil OC as a sink for CO_2 , subsoil OC attracted more attention over the last decade, with more and more reports pointing out the role of subsoil OC within the dynamic system of SOC transport and stabilization (Trumbore 2000; Baisden and Parfitt 2007; Helfrich et al. 2011). Thus, subsoil OC seems to be less inert than commonly assumed, especially with the input of fresh OC in the subsoil priming microbial processes (Fontaine et al. 2007).

Suitable for investigating subsoil OC dynamics under paddy management are the large floodplains in the coastal zone of subtropical China. Paddy and non-paddy soils (upland soils not used for submerged rice cultivation) of different ages, developed on reclaimed estuarine sediments originating from the nearby Yangtze River, were sampled near Cixi in Zhejiang Province, China. Previous results of bulk soil samples documented the transport-induced development of TOC and ¹⁴C depth gradients in the initially uniform upper 1 m of sediment, and demonstrated a replacement of old organic carbon by young OC by using the radioactive isotope ¹⁴C as a tracer (Bräuer et al. 2013). Water is the most likely transport medium, so a simple OC partition in acid/water soluble, alkali/ water soluble, and water insoluble was chosen to further evaluate OC transport. This is the classical purification scheme of ¹⁴C dating, where the insoluble, non-mobile fraction is generally chosen as the most reliable fraction for dating (e.g. Grootes et al. 2004), and closely resembles the classical separation in fulvic acids, humic acids, and humin of soil sciences. It serves to demonstrate OC transport by, and dynamic exchange of OC between, these fractions without the necessity to define specific processes and their relative importance (e.g. Kleber et al. 2007).

Elevated ¹⁴C concentrations, connected with the atmospheric testing of nuclear weapons of the late 1950s and early 1960s, make it possible to demonstrate the presence of (sub)recent plant remains in the subsoil and the potential of roots and their exudates for introducing fresh OC into deep soil horizons. Identification of rice DNA in root material at 1 m depth in a paddy soil further documents rice root penetration deep below the plough pan. We here use ¹⁴C concentrations of plant remains and of SOC fractions, characterized by different stability and mobility, to reveal the pathways of dynamic relocation of SOC into the subsoil and show that, despite the decoupling between topsoil and subsoil regarding OC accumulation (Wissing et al. 2011), subsoil OC dynamics is important for understanding the role of soils in the global carbon cycle.

METHODS

Study Sites and Soil Sampling

The study sites belong to one of the oldest paddy soil regions in the world (Cao et al. 2006). The soil samples were taken near the southern coast of Hangzhou Bay around Cixi, Zhejiang Province, China (Figure 1). With a mean annual temperature of 16.3 °C and a mean precipitation of 1325 mm per year, the climate is classified as subtropical with periodical monsoon rain (Cheng et al. 2009). Estuarine sediment, which originates from the Yangtze River, represents the parent material of the study sites. Over the past centuries, protective dikes were built to establish arable land. This led to the development of differently aged soils formed on the same substrate under comparable conditions

of climate, topography, and vegetation—the development of a soil chronosequence (Harden 1982). This circumstance allows us to study the uptake and relocation of OC with atmospheric ¹⁴C composition on comparable sites established at different times. More detailed information about the study area and the historical development of the differently aged soils is given by Cheng et al. (2009) and Kalbitz et al. (2013).



Figure 1 Geographical locations of the sampling sites at the south coast of Hangzhou Bay. To reclaim arable land, protective dikes, partly indicated by heavy black lines, were built. Years of dike construction are given next to the dike lines. Stars represent the position of sampling points (P 50, NP 50, P 100, P 1000, P 2000-yr old soil profiles, and tidal flat).

Soil horizons were described by the FAO Guidelines for Soil Description (FAO 2006) and classified according to IUSS Working Group WRB (2007). Paddy soil profiles were divided into topsoil and subsoil. Topsoils are represented by a puddle layer and plough pan (A-horizons) and obtain average depths between 20 and 25 cm. The subsoils consist of several B-horizons. More detailed information regarding the soil profile description is given by Wissing et al. (2011). In addition to paddy (P) and non-paddy (NP) soils with different ages (50 and 2000 yr), tidal flat samples (estuarine sediment) were taken. The latter represent the parent material from which soils developed in this region.

Sample Preparation

All soil samples were air-dried and sieved through 2-mm mesh size to homogenize the material. Nonsoil-derived particles as well as identifiable plant residues were removed and kept for further studies.

Chemical Extraction

Soil samples were treated with 4 mL dilute (1%) hydrochloric acid (HCl) per gram of soil to remove carbonates, and freeze-dried without washing to determine their TOC content and TOC ¹⁴C concentration. Furthermore, tidal flat, 50- and 2000-yr-old soil samples underwent chemical fractionation by acid-alkali-acid (AAA) extraction. AAA in air is a standard method in ¹⁴C dating to remove contaminating carbon from organic samples (charcoal, plant remains, etc.) prior to ¹⁴C measurements (Grootes et al. 2004). The application of AAA is similar to the classical humus fractionation procedure according to Stevenson (1994), resulting in 3 fractions: the acid-soluble fulvic acid (FA); the alkali-soluble humic acid (HA); and the insoluble humin fraction (alkali residue).

Dried soil samples were treated with diluted (1%) HCl for >10 hr to remove carbonates. The yellowish supernatant, containing acid-soluble compounds (FA), was separated. The pellet was washed with Milli-QTM water until pH > 4 to remove the hydrochloric acid and dissolved contaminants. Samples were then extracted with 1% NaOH for 4 hr at 60 °C. This yields an alkali-soluble fraction

and a non-soluble residue (humin). Humic acids were precipitated from the alkaline solution by acidification with 37% HCl down to pH < 1 and washed until pH > 4 as above. In several cases, the supernatant was kept to recover acid-soluble FAs. The insoluble residue (humin fraction) was washed with Milli-Q water until pH < 10. To remove any atmospheric CO₂ introduced during the alkali treatment, the humin fraction was again treated with 1% HCl and washed until pH > 4. All extracts, fulvic acids, humic acids, and humin fractions were freeze-dried.

Macrofossils (Plant Remains, Roots)

To obtain identifiable macrofossils, water was added to large soil samples (\sim 5 kg) and plant remains were separated by elutriation and sieving (400-µm mesh size). The sieved material was inspected under a microscope and non-plant components (stones, shells, etc.) were removed. Selected plant samples were subjected to the AAA treatment described above.

Isolation of Genomic DNA and 18S rDNA Analysis

To test whether rice plant roots reach below the plough pan, root material found at depth was tested for rice DNA by Dr Nancy Weiland-Bräuer at the Institute of Microbiology, Christian-Albrechts Universität zu Kiel (CAU). Twenty samples of root material were used for genomic DNA extraction according to Doyle and Doyle (1987) with modifications. Plant parts (300 mg) were resuspended in CTAB buffer (140 mM sorbitol, 220 mM Tris, 22 mM EDTA, 800 mM NaCl, 1% sarkosyl, 0.8% CTAB, pH 8.0)² and glass beads of 0.1-, 0.5-, and 2.5-mm size were added prior to the mechanical cell disruption for 9 min at 1300 strokes/min (GenoGrinder[®]; BT&C/OPS Diagnostics, Bridgewater, USA). Chloroform extraction of the supernatant was performed followed by precipitation of the nucleic acids with isopropanol (0.7 vol) for 20 min at room temperature and subsequent centrifugation for 1 hr at 16,000 × g and 4 °C. The DNA precipitate was washed twice with 70% ethanol and resuspended in 15 µL Tris-EDTA buffer (10 mM Tris, 1 mM EDTA, pH 8.0).

The 18S rDNA gene fragments were amplified in a standard PCR (polymerase chain reaction) using 50 ng isolated genomic DNA, GoTaq[®] polymerase (Promega, Mannheim/Germany) and primer set 18S_FW (5'-AGGAATTGACGGAAGGGCAC-3') and 18S_RV (5'-GGACATCTAAGGGCAT-CACA-3') (Luan et al. 2005). Reaction mixtures were incubated in a PT-100 thermal cycler[®] (MJ Research Inc., Waltham, USA) at 94 °C for 5 min, followed by 30 cycles at 94 °C for 30 s, 50 °C for 45 s, and 72 °C for 30 s, resulting in a 324 bp PCR fragment. DNA sequences of PCR products were determined completely by the sequencing facility at the Institute of Clinical Molecular Biology, CAU (IKM). Sequences were compared to available databases using the BLAST (Basic Local Alignment Search Tool) network service (Altschul et al. 1990) to determine their approximate phylogenetic affiliations.

Sample Treatment for AMS

After freeze-drying, sample aliquots were transferred into precombusted (4 hr, 900 °C) quartz tubes, evacuated, subsequently flame-sealed, and combusted with CuO (450 mg) and silver wool (150 mg) at 900 °C for 4 hr. The resulting CO₂ was reduced to graphite with H₂ at 600 °C over an iron catalyst. The ¹⁴C measurements were made with the 3MV HVE Tandetron AMS (accelerator mass spectrometry) system at the Leibniz-Laboratory in Kiel (Germany), with a 1 σ precision of ~0.25 pMC (Nadeau et al. 1997, 1998). ¹⁴C values are stated in percent modern carbon (pMC) and were calculated from the measured ¹⁴C/¹²C ratios of the sample and the oxalic acid standard according to Stuiver and Polach (1977).

²Tris: tris(hydroxymethyl)aminomethane; EDTA: ethylenediaminetetraacetic acid; CTAB: cetrimonium bromide.

Table 1 Organic carbon content and radiocarbonsites with different times of duration (50 yr; 200	concentration given as 0 yr; recent tidal flat)	s percent modern carb and different types of	on (pMC) of bulk soil : f land use (Paddy = lor	samples and SOM fraction wland rice cultivation	tions from 4 sampling Non-Paddy = upland
crops; Tidal Flat).					
	Bulk soil	Humin fraction	Humic acid	Fulvic acid (FA I)	Fulvic acid (FA II)

rops; lidal F	flat).																	
				Bu	ılk soil		Humi	n fract	ion	μu	nic aci	p	Fulvic	acid (F	(I Å	Fulvic a	cid (F ₁	A II)
	Sample	Depth		14C		$\mathrm{C}_{\mathrm{org}}$	¹⁴ C		$\mathrm{C}_{\mathrm{org}}$	¹⁴ C		$\mathrm{C}_{\mathrm{org}}$	¹⁴ C		$\mathbf{C}_{\mathrm{org}}$	¹⁴ C		C_{org}
Sample ID	site	(cm)	Horizon	(pMC)	+1	(%)	(pMC)	+1	(%)	(pMC)	+1	(%)	(pMC)	+1	(%)	(pMC)	+1	(%)
KIA 41465	Paddy 50	3.5	Alp	101.99	0.43	1.20	88.72	0.27	0.43	102.71	0.29	57.30	N/A	N/A	N/A	103.99	0.31	1.89
KIA 39160		12.5	Arp	101.48	0.25	1.13	83.72	0.25	0.31	102.71	0.26	24.86	N/A	N/A	N/A	106.17	0.36	0.52
KIA 39161		15	Ardp	92.36	0.25	0.49	87.61	0.31	0.17	98.38	0.23	25.36	N/A	N/A	N/A	101.54	0.57	0.12
KIA 39162		22	Ardp	57.92	0.25	0.19	34.72	0.22	0.09	68.17	0.22	10.18	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39163		31	Bwg1	49.81	0.25	0.17	22.42	0.18	0.07	64.06	0.23	8.31	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39164		39.5	Bwg2	46.98	0.25	0.19	22.60	0.19	0.08	66.64	0.24	8.00	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39165		51.5	Bwg3	48.87	0.25	0.15	24.34	0.23	0.07	63.50	0.28	9.48	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39166		71.5	Blg	50.48	0.25	0.13	28.71	0.21	0.07	63.55	0.31	9.05	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39167		91.5	Blg	49.42	0.25	0.15	30.70	0.21	0.09	65.26	0.30	9.85	N/A	N/A	N/A	N/A	N/A	N/A
KIA 41466	Non-Paddy 50	4.5	Ap	101.07	0.25	0.67	85.30	0.28	0.22	103.20	0.29	47.39	104.72	0.56	0.21	109.32	0.30	1.19
KIA 41467		13	ABw	102.06	0.25	0.54	60.23	0.28	0.11	101.85	0.34	36.52	N/A	N/A	N/A	109.42	0.30	0.97
KIA 39481		22.5	Bw	82.86	0.25	0.27	53.79	0.23	0.10	96.14	0.32	9.34	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39482		43.5	BCg	71.39	0.25	0.18	44.26	0.54	0.08	85.30	0.35	13.50	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39483		63.5	CBg	67.29	0.25	0.15	36.95	0.24	0.08	80.97	0.28	12.62	90.22	0.70	0.08	N/A	N/A	N/A
KIA 39484		81.5	CBlg	47.30	0.25	0.15	22.61	0.21	0.09	67.08	0.30	12.04	75.01	0.93	0.05	N/A	N/A	N/A
KIA 39485		101.5	CBlg	45.75	0.25	0.16	29.58	0.23	0.07	65.82	0.26	10.04	71.00	1.18	0.05	N/A	N/A	N/A
KIA 39171	Paddy 2000	13.5	Alp	112.25	0.25	3.30	116.42	0.32	1.35	111.70	0.29	32.14	112.14	0.43	3.79	114.18	0.45	1.85
KIA 39986	ı	16	Ar(d)p	113.67	0.25	3.01	116.01	0.31	1.46	111.96	0.35	8.32	113.48	0.38	2.70	114.61	0.29	1.55
KIA 39987		19	Ar(d)p	107.74	0.25	1.51	106.82	0.37	0.75	106.73	0.38	13.66	112.23	0.36	1.47	109.49	0.31	0.94
KIA 39988		26	Bdg	87.32	0.25	0.40	67.05	0.34	0.09	90.14	0.32	8.73	107.11	0.66	0.26	N/A	N/A	N/A
KIA 39989		28	2Ahgb	81.73	0.25	0.43	67.24	0.26	0.10	82.14	0.32	6.33	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39990		34	2Ahgb	73.54	0.25	0.35	52.32	0.24	0.10	74.21	0.35	6.38	68.09	0.44	0.21	N/A	N/A	N/A
KIA 39991		41.5	2Bg1	60.18	0.25	0.24	35.59	0.21	0.09	63.77	0.34	6.63	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39992		51.5	2Bg2	55.85	0.25	0.20	28.03	0.20	0.08	58.43	0.27	6.06	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39993		71.5	2Blg	36.70	0.25	0.13	18.74	0.23	0.06	51.35	0.28	5.64	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39994		91.5	2Blg	29.08	0.25	0.06	15.20	0.25	0.05	44.87	0.57	2.73	N/A	N/A	N/A	N/A	N/A	N/A
KIA 39997	Tidal flat	1	N/A	61.78	0.25	0.33	55.12	0.21	0.10	76.62	0.24	19.49	N/A	N/A	N/A	84.12	0.31	0.25
KIA 39998		19	N/A	61.32	0.25	0.31	57.00	0.22	0.12	76.69	0.25	22.94	N/A	N/A	N/A	83.29	0.33	0.17
KIA 39913		117.5	N/A	48.74	0.25	0.18	47.66	0.32	0.12	76.05	0.30	11.61	75.01	1.03	0.04	N/A	N/A	N/A

RESULTS AND DISCUSSION

Depth Gradients of ¹⁴C in Chemical Soil Organic Matter (SOM) Fractions

The ¹⁴C concentrations of TOC (Bräuer et al. 2013) and the 3 corresponding chemical fractions humin, humic acid, and fulvic acid are given as a function of depth for the 50-yr-old paddy (P50), non-paddy soil (NP50), and the 2000-yr-old paddy soil (P2000) in Figure 2a-c and Table 1, together with the values obtained for the sediment of the tidal flat. The results clearly show the age inhomogeneity of the TOC, with the insoluble humin fraction consistently older, and the more mobile humic acids younger than the TOC. The measured fulvic acids have, with 1 exception, a higher ¹⁴C concentration than the humic acids. In P50, the striking drop in ¹⁴C concentration across the plough pan (horizon Ardp), already observed for TOC (Bräuer et al. 2013), is also seen in the chemical fractions (Figure 2a). The development of their ¹⁴C concentrations with depth is very similar. The NP50 shows a more or less constant decrease in ¹⁴C with depth for all SOC fractions, comparable to ¹⁴C gradients reported by Scharpenseel et al. (1989) and Paul et al. (1997). Thus, the decoupling of topsoil from subsoil by plough pan development under rice paddy management clearly affects all SOC fractions in the subsoil and delays the buildup of the customary ¹⁴C or "age"-depth profile beyond 50 yr. After 2000 yr, such a profile is also found in the paddy subsoil (Figure 2c).



Figure 2 ¹⁴C concentrations of bulk soil samples and the corresponding SOM fractions (humin, humic acid, fulvic acid) for P50, NP50, and P2000. The stars display ¹⁴C concentrations of the parent material, represented by tidal flat samples from depths of 0–2, 18–20, and 115–120 cm (black: humin fraction; dark gray: bulk soil; light gray: humic acid; white: fulvic acid). ¹⁴C measurement uncertainties are smaller than the symbol sizes.

After 50 yr of cultivation, bulk soil samples from the topsoil of both the paddy and non-paddy gave ¹⁴C contents above 100 pMC, close to the atmospheric ¹⁴CO₂ level (105 pMC) in 2008, the year of sampling (Levin et al. 2010; Graven et al. 2012). This indicates that, already after 50 yr, a high proportion of the original OC of the tidal flat sediments, with ~61 pMC, had been replaced with OC

derived from recent atmospheric photosynthesis products. Both P50 and NP50 topsoils show comparable ¹⁴C concentrations for all SOC fractions. The humin fraction, with ¹⁴C values around 85 pMC, still clearly shows the influence of the original OC. The rapid ¹⁴C increase in this relatively stable fraction may be due more to the increase in TOC and humin OC in the paddy topsoil than to replacement. All topsoil fractions of P2000 are characterized by ¹⁴C values significantly higher than those of the P50 and NP50 sites (Table 1). The high ¹⁴C and TOC values can be assigned to the accumulation of plant-derived carbon during the last several decades. Interestingly, the humin fraction, which is assumed to be the most stable part of SOC, shows the highest modern ¹⁴C concentrations within the uppermost 15 cm (116.4 pMC in horizon Alp). These values, which exceed the atmospheric ¹⁴C level in 2008, indicate the contribution of OC from the last 40 to 50 yr (average ¹⁴CO₂ = 116.6 pMC in 1989). The ¹⁴C concentrations of the humic and fulvic acids in the P2000 topsoil, (slightly) below those of the TOC though still well above the atmosphere of 2008, suggest these derive on average from more recent decomposition of organic material.

The observed decrease in 14 C in the humin fraction with increasing soil depth is commonly observed and has been interpreted in previous reports by Balesdent (1987), Pessenda et al. (2001), and Rethemeyer et al. (2005) as aging and a relative increase of recalcitrant organic compounds in the humin fraction (Rice 2001). Yet, in the case of this chronosequence, it documents the replacement of the original estuarine OC by younger components moving down the soil profile, presumably as DOC, and then becoming incorporated in the humin fraction. Right below the P2000 plough pan with 116 pMC, the humin fraction shows 67 pMC, already indicating mostly aged, recalcitrant SOC with an apparent age of ~3000 14 C yr, yet almost 20 pMC enriched in 14 C compared to the original estuarine sediments of the tidal flat.

The humin fraction shows ¹⁴C contents lower than those of the initial tidal flat sediment at depth in all profiles (Figure 2). The decrease is too large to be caused by natural decay, even in 2000 yr, which would reduce the 47.6 pMC of the deep tidal flat to 37.4 pMC. Instead, the low ¹⁴C concentrations of the deep humin fractions may reflect the effect of a supply of fresh OC to a subsoil containing a mixture of organic material, in part old, reworked, and in part contemporaneous with the time of sedimentation (Bräuer et al. 2013). With measured ¹⁴C concentrations in the deep tidal flat close to 50 pMC, the contemporaneous fraction must be at least 50% of the bulk and humin OC. The values close to 23 pMC at 30 to 40 cm in P50, and around 80 cm in NP50, as well as the values below 20 pMC in P2000 below 70 cm (Table 1), indicate only 23% or less "young" OC remained in these humin fractions, which means that 65% or more of the younger components of the estuarine OC mixture were preferentially degraded and lost. Interestingly, in P50, where the ploughpan still has blocked most of the downward transport of OC, the minimum ¹⁴C concentration is found closely below this ploughpan, while in NP50 and P2000 minimum concentrations are found at depth, near the groundwater table. This suggests a relatively rapid preferential degradation of younger OC stimulated by a supply of fresh OC ("priming") and a gradual buildup of more recalcitrant young OC. The ~50 pMC still found at depth in the tidal flat may then reflect absence of vegetation and anoxic conditions.

The extracted FAs differ in ¹⁴C concentration from the humic acids and the humin, and show small, yet significant differences before and after alkali treatment. In the P2000 topsoil, the ¹⁴C of all FAs is above that of the 2008 atmosphere (105 pMC, Levin et al. 2010; Graven et al. 2012); in NP50 this is only true for FA II, while in P50 only one FA II exceeds the atmosphere. With the exception of the Bdg horizon at 26 cm in P2000, the FAs in the subsoil have a ¹⁴C concentration below 100 pMC. In the P2000 topsoil, the ¹⁴C in FA II is higher than in FA I at 13.5 and 16 cm depth (Table 1) and closer to the humin fraction from which it was separated. At 19 cm, FA II has less ¹⁴C than FA I and, thus,

is again closer to the humin fraction. This apparent link between FA II and the humin fraction seems broken in the NP50 topsoil, where FA II shows a ¹⁴C concentration well above that of the 2008 atmosphere, although all other fractions are below it. This seemingly erratic pattern may be ascribed to the diverse components of the organic mixture. The starting material was a mixture of old, reworked OC and young OC with a ¹⁴C concentration, presumably, close to 100 pMC, to which plant material was added over the years with ¹⁴C concentrations increasing irregularly from 98 pMC in 1954 to ~180 pMC in 1964 and then gradually dropping to 105 pMC in 2008 (Levin et al. 2010; Graven et al. 2012). This is reflected in the high pMC values of the humin fraction in the P2000 topsoil, indicating OC on average a few decades old with high bomb-¹⁴C concentrations. The FA I fraction is fairly small (Figure 3) and contains readily accessible and exchangeable components (Kleber et al. 2007) with a ¹⁴C closer to the recent atmosphere.



Figure 3 The proportions of single SOM fractions on the total organic carbon (TOC) given in percent. Quantification of humic acids and humin fraction yields fulvic acids (FA) as a difference with TOC. The FA of samples without a stated loss was calculated from the TOC, humin, and HA fraction amounts and not isolated. The mass balance for samples, from which FAs were extracted for ¹⁴C measurement, is incomplete due to material loss during wet chemistry sample preparation. In cases where both FA fractions, before (FA I) and after alkali treatment (FA II), were recovered, FA II always delivers proportions on TOC at least 2 times (maximum $6\times$) higher than FA I.

Relative Importance of Mobile Fractions

The relative contributions of the SOC fractions to the TOC are given in Figure 3. As described in the Methods section, the humin fraction is what remains after AAA extraction. The humic acid fraction is precipitated from the alkali extract with acid and washed until free of acid and soluble organic compounds. Quantification of the humic acids and the humins generally yields calculated fulvic acid (FA) contents as the difference with TOC (Figure 3). Direct recovery of the complete fulvic acid fraction is difficult, since it requires collection of all water used in the various washings. In a number of cases, fulvic acids were recovered by freeze-drying the first acid extract (FA I) and the acid supernatant after precipitation of the humic acid fraction from the alkali extract (FA II). For practical reasons, the subsequent, ever more diluted, washings were not included. The isolated ful-

vic acids thus represent only a fraction of the TOC not accounted for by the humin and humic acid fractions (10 to 50% for FA I plus FA II, Figure 3). The missing material was assigned to the fulvic acid fraction, although it may include some fine colloidal material, which chemically does not belong to the fulvic acids. Interestingly, FA II generally contains more OC than FA I. While slightly acidic water can mobilize only a small fraction of the available SOC, as generally observed, the 4 hr of alkaline extraction at 60 °C apparently loosens chemically bound compounds (e.g. Kleber et al. 2007), many of them belonging to the fulvic acid fraction, which were resisting the acidic solution.

With an average contribution of close to 50% of the TOC, the so-defined FAs constitute the main SOC fraction in each profile. The average contribution of the humic acid fraction is <20% for all soil profiles, while the humin fraction represents ~30% of the TOC. The importance of the insoluble humin fraction increases with increasing depth (Figure 3), up to 47% of TOC at depth in P50 and 67% in P2000, but only 36% in NP50. Since both TOC and the humin fraction decrease strongly with depth, especially from plough pan to subsoil (e.g. Table 1, TOC from ~1.2% C in the P50 topsoil to 0.14% C at depth, from 3.3% C in the P2000 topsoil to ~0.1% at depth), this relative increase reflects a very strong decrease in the more mobile fractions, especially in the FAs, in the subsoil. The TOC in the deeper subsoil is more or less the same in all 3 profiles (P50, NP50, and P2000) and similar to, or slightly lower than, that in the deep tidal flat. The same is true for the 3 humin fractions, which are, however, reduced relative to the tidal flat. The carbon mass balance thus provides the picture of strongly decreasing mobile components and little change in the subsoil that led to the neglect of subsoil processes.

The buildup of ¹⁴C concentration profiles in the originally homogeneous, ¹⁴C-depleted sediment of the subsoil documents, however, the downward transport of SOC, presumably as dissolved organic carbon (DOC), with infiltrating water. FAs are of major importance for this transport of OC through the soil profile as indicated by their high proportion, especially within the transition zone from plough pan to subsoil. Maie et al. (2004) reported chemical characteristics of FAs leached from the plough layer of a paddy soil, where the FA fraction was up to 56% of the OC. The liberation of FAs by alkali extraction and the parallel behavior of the ¹⁴C-depth curves for humins, humic acids, and fulvic acids (Figure 2) suggest a dynamic exchange between these 3 SOC fractions and a likely role of the mobile FA and HA fractions as main driver for the described refreshment of SOC pools below plough pan level by organic matter with a younger ¹⁴C signature.

Plant Remains: Another Pathway of OC Transport into Subsoil

In addition to visual methods, molecular biological methods were applied to identify plant remains found deep in soil profiles. This is particularly useful when only parts of plants are available. The characterization is based on isolation of genomic plant DNA and subsequent amplification of the 18S rDNA fragment with specific primers in a PCR. Plant remains recovered by wet-sieving from all depths of the soil profiles were differentiated by morphological means. Most of these plant-derived macrofossils were found in the topsoil of each profile. The assemblage consists of seeds, spelts, and grains of several species like millet, wheat, barley, or rice. Furthermore, straw, rice hulls, roots, and charcoals were found. Collecting such plant remains below the plough pan was more difficult due to their rare occurrence. Most of the plant remains and root mats. Thus, classification into a species by visual means was not possible. DNA extraction and analysis identified 1 out of 20 root samples as rice (*Oryza sativa*), indicating rice root penetration below the plough pan to a surprising depth of 97–100 cm (Table 2).

Rice is often described as a shallow-rooting crop with roots mostly located above the plough pan, which restricts deep-rooting growth (Wade et al. 1999). Yet, Mishra et al. (1997) reported a maximum rooting depth between 55 and 65 cm, depending on water table conditions. The rooting depth is assumed to be influenced by the prevalent water regime, so dry-land varieties are characterized by deeper rooting depths, while submerged rice plant varieties tend to be shallow-rooted (Yoshida and Hasegawa 1982). Soil cracks that develop when the soil dries out facilitate root penetration below the plough pan. Thus, the presence of roots, including those of rice, is not totally unexpected and indicates that young carbon may be inserted directly into the subsoil as root material and root exudates.

Table 2 Radiocarbon and organic carbon contents of plant remains (roots and root mats) selected from deep subsoil horizons of different aged paddy soils. The table includes also ¹⁴C data of plant remains of 100- and 1000-yr-old paddy soils. No data are available for the P2000 soil profile. The presented data belong to the samples that were used for DNA analysis.

			Sampled depth	¹⁴ C	Corg
Sample ID	Sample site	Horizon	(cm)	(pMC)	(%)
KIA 39164	Paddy 50	Bwg2	38–41	112.08 ± 0.47	52.34
KIA 39166	Paddy 50	Blg	70–73	119.57 ± 1.03	48.44
KIA 40291	Paddy 100	Bwlg1	72–75	110.08 ± 0.46	37.11
KIA 40292 ^a	Paddy 100	Bwlg2	97-100	127.57 ± 0.75	41.88
KIA 40324	Paddy 1000	3B1	107–110	118.34 ± 0.33	49.28

^aThis is the only sample that could be identified as rice (Oryza sativa).

Although the roots formed a root mat, it cannot be excluded that plant and root remains were relocated down by bioturbation, since canals and holes of up to several mm diameter as well as living earthworms as indicators for bioturbation activity were observed during sampling across the entire soil profile. Recent studies identified such macropore systems as flow paths, and therefore revealed crack and root channels as well as earthworm burrows as responsible for water loss and solute leaching (Janssen and Lennartz 2008; Lennartz et al. 2009).

CONCLUSIONS

Soil profiles from a chronosequence of plots under paddy and non-paddy management formed on the same uniform ¹⁴C-depleted estuarine sediments reveal the rapid development of ¹⁴C and TOC gradients in the subsoil as a result of organic carbon mobility and relocation. The following are the processes of subsoil "OC refreshing" indicated by ¹⁴C concentrations:

- 1. Direct input of fresh (modern) SOC into deep subsoil by plant roots and their exudates. This process is evident by roots and root remains, including rice roots identified by their DNA signature, with modern ¹⁴C signature found far below the plough pan in deep soil horizons.
- 2. Relocation and transport of young OC through the plough pan via mobile SOC fractions (humic acids and fulvic acids). Due to their high proportion on total organic carbon, we assume fulvic acids to be the main driver for the input of OC through the plough pan into the subsoil.
- 3. Downward movement of plant remains into the subsoil through animal activity and soil cracking.

Despite this input of fresh organic material, OC concentrations in the subsoil remain largely constant over the chronosequence, while the ¹⁴C concentration increases. The composition of subsoil OC down to the 1 m depth of this investigation is thus determined by a dynamic equilibrium between the import of fresh OC, mineralization, and export into groundwater and deeper subsoil.

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REFERENCES

- Altschul S, Gish W, Miller W, Myers E, Lipman D. 1990. Basic local alignment search tool. *Journal of Molecular Biology* 215(3):403–10.
- Baisden WT, Parfitt RL. 2007. Bomb ¹⁴C enrichment indicates decadal C pool in deep soil? *Biogeochemistry* 85(1):59–68.
- Baisden WT, Amundson R, Brenner DL, Cook AC, Kendall C, Harden JW. 2002. A multiisotope C and N modeling analysis of soil organic matter turnover and transport as a function of soil depth in a California annual grassland soil chronosequence. *Global Biochemical Cycles* 16(4):8210–26.
- Balesdent J. 1987. The turnover of soil organic fractions estimated by radiocarbon dating. *The Science of the Total Environment* 62:405–8.
- Batjes NH. 1996. Total carbon and nitrogen in the soils of the world. *European Journal of Soil Science* 47: 151–63.
- Becker-Heidmann P, Scharpenseel HW. 1992. Studies of soil organic matter dynamics using natural carbon isotopes. *The Science of the Total Environment* 117/118: 305–12.
- Bräuer T, Grootes PM, Nadeau MJ, Andersen N. 2013. Downward carbon transport in a 2000-year rice paddy soil chronosequence traced by radiocarbon measurements. *Nuclear Instruments and Methods in Physics Research B* 294:584–7.
- Cao ZH, Ding JL, Hu ZY, Knicker H, Kögel-Knabner I, Yang LZ, Yin R, Lin XG, Dong YH. 2006. Ancient paddy soils from the Neolithic age in China's Yangtze River Delta. *Naturwissenschaften* 93(5):232–6.
- Cheng YQ, Yang LZ, Cao ZH, Yin S. 2009. Chronosequential changes of selected pedogenic properties in paddy soils as compared with non-paddy soils. *Geoderma* 151:31–41.
- Christensen BT. 2001. Physical fractionation of soil and structural and functional complexity in organic matter turnover. *European Journal of Soil Science* 52:345– 53.
- Doyle JJ, Doyle JL. 1987. A rapid DNA isolation procedure for small quantities of fresh leaf tissue. *Phytochemistry Bulletin* 19:11–5.
- Dreves A, Andersen N, Grootes PM, Nadeau M-J, Garbe-Schönberg CD. 2007. Colloidal matter in water extracts from forest soils. *Environmental Chemistry*

4(6):424-9.

- FAO. 2006. *Guidelines for Soil Description*. Rome: FAO. 97 p.
- FAOSTAT (Food and Agricultural Organization of the United Nations). 2010. URL: http://faostat.fao.org/.
- Flessa H, Amelung W, Helfrich M, Wiesenberg GLB, Gleixner G, Brodowski S, Rethemeyer J, Kramer C, Grootes PM. 2008. Storage and stability of organic matter and fossil carbon in a Luvisol and Phaeozem with continuous maize cropping: a synthesis. *Journal* of Plant Nutrition & Soil Science 171:36–51.
- Fontaine S, Barot S, Barre P, Bdioui N, Mary B, Rumpel C. 2007. Stability of organic carbon in deep soil layers controlled by fresh carbon supply. *Nature* 450(7167): 277–80.
- Graven HD, Guilderson TP, Keeling RF. 2012. Observations of radiocarbon in CO₂ at seven global sampling sites in the Scripps flask network: analysis of spatial gradients and seasonal cycles. *Journal of Geophysical Research* 117: D02303, doi:10.1029/2011JD016535.
- Grootes PM, Nadeau MJ, Rieck A. 2004. ¹⁴C-AMS at the Leibniz-Labor: radiometric dating and isotope research. *Nuclear Instruments and Methods in Physics Research B* 223–224:55–61.
- Harden JW. 1982. A quantitative index of soil development from field descriptions: examples from a chronosequence in central California. *Geoderma* 28:1–28.
- Helfrich M, Flessa H, Ludwig B. 2010. Modelling carbon dynamics in subsoils using simple models. *Journal of Plant Nutrition and Soil Science* 173(5):671–7.
- Huang Y, Sun W. 2006. Changes in topsoil organic carbon of croplands in mainland China over the last two decades. *Chinese Science Bulletin* 51(15):1785–803.
- IUSS Working Group WRB. 2007. World Reference Base for Soil Resources. World Soil Resources Reports, Volume 103. Rome: FAO. 145 p.
- Janssen M, Lennartz B. 2008. Characterization of preferential flow pathways through paddy bunds with dye tracer tests. Soil Science Society of America Journal 72(6):1756–66.
- Jobbagy EG, Jackson RB. 2000. The vertical distribution of soil organic carbon and its relation to climate and vegetation. *Ecological Applications* 10(2):423–36.
- Kalbitz K, Kaiser K, Fiedler S, Kölbl A, Amelung W, Bräuer T, Cao ZH, Don A, Grootes PM, Jahn R,

Schwark L, Vogelsang V, Wissing L, Kögel-Knabner I. 2013. The carbon count of 2,000 years of rice cultivation. *Global Change Biology* 19(4):1107–13.

- Kleber M, Sollins P, Sutton R. 2007. A conceptual model of organo-mineral interactions in soils: self-assembly of organic molecular fragments into zonal structures on mineral surfaces. *Biogeochemistry* 85:9–24.
- Kögel-Knabner I, Amelung W, Cao ZH, Fiedler S, Frenzel P, Jahn R, Kalbitz K, Kölbl A, Schloter M. 2010. Biogeochemistry of paddy soils. *Geoderma* 157:1–14.
- Lal R. 2004. Soil carbon sequestration impacts on global climate change and food security. *Science* 304(5677): 1623–6.
- Lal R. 2008. Carbon sequestration. *Philosophical Trans*actions of the Royal Society B 363:815–30.
- Laskar AH, Yadava MG, Ramesh R. 2012. Radiocarbon and stable isotopes in two soil profiles from northeast India. *Radiocarbon* 54(1):81–9.
- Lennartz B, Horn R, Duttmann R, Gerke HH, Tippkötter R, Eickhorst T, Janssen I, Janssen M, Rüth B, Sander T, Shi X, Sumfleth K, Taubner H, Zhang B. 2009. Ecological safe management of terraced rice paddy landscapes. Soil & Tillage Research 102:179–92.
- Levin I, Naegler T, Kromer B, Diehl M, Francey RJ, Gomez-Pelaez AJ, Steele LP, Wagenbach D, Weller R, Worthy DE. 2010. Observations and modeling of the global distribution and long-term trend of atmospheric ¹⁴CO₂. *Tellus B* 62:26–46.
- Luan YX, Mallatt JM, Xie RD, Yang YM, Yin WY. 2005. The phylogenetic positions of three basal-hexapod groups (Protura, Diplura, and Collembola) based on ribosomal RNA gene sequences. *Molecular Biology* and Evolution 22:1579–92.
- MacLean JL, Dawe DC, Hardy B, Hettel GP. 2002. Rice Almanac: Source Book for the Most Important Economic Activity on Earth. Wallingford: CABI Publishing. 270 p.
- Maie N, Watanabe A, Kimura M. 2004. Chemical characteristics and potential source of fulvic acids leached from the plow layer of paddy soil. *Geoderma* 120: 309–23.
- Marschner B, Brodowski S, Dreves A, Gleixner G, Gude A, Grootes PM, Hamer U, Heim A, Jandl G, Ji R, Kaiser K, Kalbitz K, Kramer C, Leinweber P, Rethemeyer J, Schaeffer A, Schmidt MWI, Schwark L, Wiesenberg GLB. 2008. How relevant is recalcitrance for the stabilization of organic matter in soils? *Journal of Plant Nutrition & Soil Science* 171(1):91–110.
- Mishra HS, Rathore TR, Pant RC. 1997. Root growth, water potential, and yield of irrigated rice. *Irrigation Science* 17:69–75.
- Nadeau MJ, Schleicher M, Grootes PM, Erlenkeuser H, Gottdang A, Mous DJW, Sarnthein JM, Willkomm H. 1997. The Leibniz-Labor AMS facility at the Christian-Albrechts University, Kiel, Germany. *Nuclear In*struments and Methods in Physics Research B 123(1– 4):22–30.

- Nadeau MJ, Grootes PM, Schleicher M, Hasselberg P, Rieck A, Bitterling M. 1998. Sample throughput and data quality at the Leibniz-Labor AMS facility. *Radiocarbon* 40(1):239–45.
- O'Brien BJ, Stout JD. 1978. Movement and turnover of soil organic matter as indicated by carbon isotope measurements. *Soil Biology and Biochemistry* 10(4): 309–17.
- Pan G, Li L, Wu L, Zhang X. 2003. Storage and sequestration potential of topsoil organic carbon in China's paddy soils. *Global Change Biology* 10:79–92.
- Paul EA, Follett RF, Leavitt SW, Halvorson A, Petersen GA, Lyon DJ. 1997. Radiocarbon dating for determination of soil organic matter pool sizes and dynamics. *Soil Science Society of America Journal* 61:1058–67.
- Pessenda LCR, Gouveia SEM, Aravena R. 2001. Radiocarbon dating of total soil organic matter and humin fraction and its comparison with ¹⁴C ages of fossil charcoal. *Radiocarbon* 43(2B):595–601.
- Rethemeyer J, Kramer C, Gleixner G, John B, Yamashita T, Flessa H, Andersen N, Nadeau MJ, Grootes PM. 2005. Transformation of organic matter in agricultural soils: radiocarbon concentration versus soil depth. *Geoderma* 128(1–2):94–105.
- Rice JA. 2001. Humin. Soil Science 166(11):848-57.
- Rui W, Zhang W. 2010. Effect size and duration of recommended management practices on carbon sequestration in paddy field in Yangtze Delta Plain of China: a meta-analysis. Agriculture, Ecosystems and Environment 135:199–205.
- Rumpel C, Kögel-Knabner I. 2011. Deep soil organic matter – a key but poorly understood component of terrestrial C cycle. *Plant and Soil* 338(1–2):143–58.
- Scharpenseel HW, Becker-Heidmann P, Neue HU, Tsutsuki K. 1989. Bomb-carbon, ¹⁴C-dating and ¹³Cmeasurements as tracers of organic matter dynamics as well as of morphogenetic and turbation processes. *The Science of the Total Environment* 81/82:99–110.
- Six J, Guggenberger G, Paustian K, Haumaier L, Elliott ET, Zech W. 2001. Sources and composition of soil organic matter fractions between and within soil aggregates. *European Journal of Soil Science* 52:607–18.
- Stevenson FJ. 1994. Humus Chemistry Genesis, composition, reactions. New York: John Wiley. 512 p.
- Stuiver M, Polach HA. 1977. Discussion: reporting of ¹⁴C data. *Radiocarbon* 19(3):355–63.
- Tanji KK, Gao S, Scardaci SC, Chow AT. 2003. Characterizing redox status of paddy soils with incorporated rice straw. *Geoderma* 114(3–4):333–53.
- Trumbore SE. 2000. Age of soil organic matter and soil respiration: radiocarbon constraints on belowground C dynamics. *Ecological Applications* 10(2):399–411.
- Trumbore SE, Vogel JS, Southon JR. 1989. AMS ¹⁴C measurements of fractionated soil organic matter: an approach to deciphering the soil carbon cycle. *Radiocarbon* 31(3):644–54.
- Wade LJ, Fukai S, Samson BK, Ali A, Mazid MA. 1999.

Rainfed lowland rice: physical environment and cultivar requirements. *Field Crop Research* 64:3–12.

- Wang Y, Amundson R. 1996. Radiocarbon dating of soil organic matter. *Quaternary Research* 45:282–8.
- Wissing L, Kölbl A, Vogelsang V, Fu JR, Cao ZH, Kögel-Knabner I. 2011. Organic carbon accumulation in a 2000-year chronosequence of paddy soil evolution. *Catena* 87:376–85.
- Xu S, Shi X, Zhao Y, Yu D, Li C, Wang S, Tan M, Sun W. 2011. Carbon sequestration potential of recommended management practices for paddy soils of China, 1980–2050. *Geoderma* 166(1):206–13.
- Yoshida S, Hasegawa S. 1982. The rice root system: its development and function. In: *Drought Resistance in Crops with Emphasis on Rice*. Manila: IRRI. p 97–114.