Grazing Impacts in Vegetated Dune Fields: Predictions From Spatial Pattern Analysis

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INTRODUCTION

Arid and semiarid ecosystems compose over one-third of the land surface of the world (OIES 1991). Among these, vegetated dune fields are systems that are extremely fragile and sensitive to disturbances because of high susceptibility to erosion of sandy soils (UNEP 2006). Vegetation protects the surface via direct cover of the surface, trapping of particles, and most importantly by extracting momentum from the air flow (Wolfe and Nickling 1993). Vegetation cover varies with position on the dune—flanks are normally better vegetated than crests owing to soil moisture distributions and degree of exposure to wind events (Wiggs et al. 1995). There is also significant short-term variation in dune system dynamics because of high interannual rainfall variability, droughts, localized fire, and human impacts such as livestock grazing (Wasson and Nanninga 1986). These factors can reduce vegetation cover leading to aeolian (wind-borne) sediment mobilization (Living-
Livestock grazing can generate and accelerate soil erosion processes through intensive utilization of the rangeland (Pickup et al. 1994). Intensive grazing can be conditioned, among other factors, by the distribution of water points. In large grazed paddocks, with a single water point located at one corner, a direct trend predominates when gradients of several paddocks are pooled in a regional estimate: higher grazing pressure around water points, diminishing at increasing distance from it (Bastin et al. 1993). Lange (1969) coined the term piosphere for these utilization gradients.

Grazing impact is greatest close to a watering point and decreases with distance from the water for two reasons: 1) the area available to graze increases with distance from the focal point resulting in a reduction in density of stock and 2) stock have to drink regularly so they are limited in how far they can travel from water (James et al. 1999). As well as grazing effects, there are also effects from trampling associated with the movement of animals close to the watering point (Andrew and Lange 1986). Trampling is most obvious within a few hundred meters of the watering point, an area often called the ‘sacrifice zone’ (Valentine 1947).

Deterioration of arid rangelands caused by livestock can often be reverted by applying suitable methods of ecosystem management. However, traditional field-based studies on dryland ecosystems are usually confined to restricted areas or small samples, such as transects, suites of quadrats, and inside–outside-enclosure comparisons (Sparrow et al. 2003). Therefore, knowledge about the grazing process at the whole-paddock scale comes by extrapolation, sometimes supported by modeling, without considering that patterns and processes acting at broader scales may not be recognized when extrapolating from small to large areas (Ludwig et al. 2000). A different approach to studying effects of grazing on arid landscapes is achieved with spatial pattern analysis techniques that have been widely adopted recently (Ripley 1981; Haase 1995; Barot et al. 1999; Jeltsch et al. 1999; Wiegand and Moloney 2004). In accordance with this approach, analysis of spatial patterns in rangelands is used to infer the existence of underlying processes, which may operate at different spatial scales in the paddock. However, the number of studies that apply such techniques is still low in rangeland ecology (Adler et al. 2001).

The assessment of the impact of grazing on soil erosion processes prevailing in arid rangelands requires the selection of adequate indicators, or relevant process-related parameters to quantify them. Blowouts have been indicated as the most common aeolian erosional landforms in dune landscapes in arid zones (Hesp and Hyde 1996). Blowouts are depressions or hollows formed by wind erosion on a preexisting sand deposit (Hesp 2002) and are common features in vegetation-stabilized dune fields (Livingstone and Warren 1996). Vegetated parabolic dunes are leeward extensions of blowouts. Plants colonize, and may eventually stabilize, the trailing arms, which are formed by the advancing dune apex (Melton 1940; Verstappen 1968; Pye 1982; Eriksson et al. 1989). Grazing, however, can contribute to destabilization of dunes (e.g., de Stoppelaire et al. 2004). The vegetated dune field at Peninsula Valdés (northeastern coast of Patagonia) is a good system to deal with spatial patterns of soil erosion processes. Dunes were formed in the upper Pleistocene from loose, sand-sized sediment of the sandy beaches and cliffs exposed to the prevalent westerly winds (Beltramone et al. 1990). Nowadays, dunes are mostly stabilized and the presence of blowouts is the best evidence of erosive processes in the dune field (Súmico 1996).

The emphasis of this paper is on the analysis of the changes induced by grazing on soil erosion processes prevailing in a vegetated dune field at Peninsula Valdés, using the spatial pattern of blowouts as an indicator of erosion intensity. We used distance-dependent spatial statistics to identify scales at which the spatial pattern of blowouts was significantly aggregated around water points (i.e., “critical scales”). We hypothesized that large numbers of livestock concentrated at water points may destroy vegetation cover, resulting in dune reactivation and consequently generating blowouts. Accordingly, we expected that the impact of livestock on soil erosion processes was greatest in sites near water points relative to those farther from water points (i.e., a utilization gradient). This gradient should result in aggregation of blowouts near water points. However, the spatial structure of the sites sensitive to erosion (i.e., dune crests) may confound a direct effect. We therefore analyzed blowout densities at different distances away from water points to see if they were higher than expected by a random distribution of blowouts over the dune crest (i.e., after removing potential spatial structures). Besides, the critical scale of aggregation should increase under grazing. Considering that vegetation cover varies with position on the dune, we also analyzed the effects of the spatial distribution of dune crests around water points on the spatial pattern of blowouts. We hypothesized that the effect of livestock on soil erosion processes is influenced by the preexisting spatial pattern of dune crests. Thus, we expected that the grazing disturbance was greatest in sites with a higher density of dune crests.

**METHODS**

**Study Site Description**

The dune field is located in the southern portion of the Peninsula Valdés (lat 42°32’S, long 63°54’W) in the northeastern region of Patagonia, southern Argentina. The climate is semiarid, characterized by an annual mean temperature of 13°C and an average annual rainfall (1912–2002) of 231 mm, with high mean interannual variation (coefficient of variation = 30%; Barros and Rivero 1982). The prevailing winds are from the west and northwest (Coronato 1994). Mean annual wind speed is 23 km · h⁻¹ (Barros et al. 1981). It is highly influenced by intense winds from the northeast prevailing from October to February (Labraga 1994). Sheep were introduced in the area at the beginning of last century (Defosse et al. 1992) and are presently raised in large _estancias_ (10 000 ha or more), usually consisting in several paddocks around a shared water point.

The source of sediment for the dune field are the sandy beaches of Nuevo Gulf where a continued supply of loose, sand-sized sediment is available to be transported inland by the prevailing westerly winds (Haller et al. 2000). General features
in the topography of the dune field are relic aeolian landforms, megapatches of active sand dunes, and erosional features such as blowouts and deflation plains. Relict aeolian landforms would have been formed in a periglacial environment (Iriondo and García 1993; Trombotto 1998), and include sand sheets and linear dunes stabilized by psammophile species such as Sporobolus rigens and Hyalis argentea (Bertiller et al. 1981). Megapatches of active sand dunes include a series of aeolian forms as barchan, dome, and transverse dunes. The currently visible wind shaping of these dunes has been active since the upper Pleistocene (Beltrame et al. 1990).

Sizes and shapes of blowouts on the dune field at Peninsula Valdés vary considerably. Blowouts have lengths varying from just a few meters to more than 100 m, and morphologies that range from circular hollows to elongated corridors. Some of them are associated with parabolic dunes. Blanco (2004) measured the migration rates of 15 parabolic dunes for the period 1969–2002. The average distance traveled by parabolic dunes in the dune field between 1969 and 2002 was 334 m, equivalent to an average migration rate of 10.1 m · yr⁻¹. From the calculated migration rates and measured distance between the origin of blowouts and the edge of the deflation plain in 1969, it can be calculated how many years it would take the blowouts to reach this length. The maximum and lowest ages for blowouts associated with parabolic dunes were estimated at 71 and 32 yr (Blanco 2004).

Grazing Rates

We studied paddocks in which the stocking rate and the location of fences and water had remained fairly unchanged for at least the last four decades. Therefore, the present study was limited to eight paddocks (covering a total of about 6 600 ha) under two sheep grazing intensities: four paddocks were lightly grazed and four heavily grazed. We calculated that lightly grazed paddocks had a stocking rate ca. 0.4 sheep · ha⁻¹ and heavily grazed paddocks ca. 0.8 sheep · ha⁻¹. Each of the paddocks had one single water point. The maximum distance in each paddock from the water point ranged between 1.9 and 4.6 km.

Mapping Blowouts, Dune Crests, and Water Points

Blowouts and dune crests were identified using historical aerial photographs, taken in January 1969 by the Instituto Geográfico Militar. For a current perspective of these aeolian features we used a Landsat 7 Enhanced Thematic Mapper Plus image from 10 October 2002 with a spatial resolution of 30 m and with a minimum mapping unit of 9 × 10⁻² ha. Aerial photographs were scanned at a resolution of 600 dots per inch and converted to tagged interchange format (TIFF) for use in Erdas Imagine software (Erdas Inc. 2003). These images were georeferenced and projected into a Transverse Mercator projection with WGS 1984 spheroid and datum. The maximum error between the transformation model and the reference coordinates was less than one pixel. The maps were scaled at 1:60 000, with a pixel resolution of 6 m and a minimum mapping unit of 0.36 × 10⁻² ha. The georeferenced aerial photos were joined together to form a single rectified mosaic.

We mapped the polygons of erosional features (blowouts and deflation plains) on the mosaic of aerial photos. Point maps were then made for each paddock in order to represent the positions of origins of blowouts and deflation plains. Also, the crests of stabilized linear dunes were mapped and arc maps were obtained. Furthermore, we generated point maps of water point locations. All maps were exported as shapefiles.

Aerial photos were used to geometrically correct, coregister, and project the Landsat image. Stable points, such as cultural landmarks that could be located on the images, were used to coregister and rectify the imagery. A root mean square error of less than one pixel was obtained. The image was then classified using a supervised classification scheme with a maximum likelihood classifier (Erdas Inc. 2003). Three classes were defined: 1) blowouts and deflation plains, 2) active sand dunes, and 3) vegetated areas. Ground-truth data were obtained in summer 2003 to determine the classification accuracy. Raster blowouts and deflation plains were vectorized as polygons and then we built point maps representing them for each paddock. These maps were exported as shapefiles.

The geographical information systems software ArcView 3.3 (ESRI 2002) was used to transform shapefile data into a rasterized Arc/Info grid format, with a resolution of 30 × 30 m. For each paddock, a map was generated from the addition of grids of blowouts, dune crests, and water points (Fig. 1). The resulting maps were exported in ASCII 3 format.

Analyzing Blowout Spatial Patterns

Blowouts spatial patterns were analyzed with the O-ring statistic (Wiegand et al. 1999) using the Programita software (Wiegand and Moloney 2004). The bivariate O-ring statistic \( O_{12}(r) \) describes the spatial relationship between two types of points (type 1 points and type 2 points) as function of distance \( r \), summarizing the information on the distances between all pairs of type 1-type 2 points. More formally, the bivariate O-ring statistic \( O_{12}(r) \) can be defined as the density of type 2 points in a ring with width \( w \) at distance \( r \) from an arbitrary type 1 point (Wiegand and Moloney 2004). Note that the \( O_{12}(r) \) is related to the derivative of Ripley’s K function (Ripley 1981), i.e., \( O_{12}(r) = \lambda_2 K_{12}'(r)/(2\pi r) \) where \( \lambda_2 \) is the intensity of the pattern of type 2 points in the study area (Wiegand and Moloney 2004). The main difference between \( O_{12} \) and \( K_{12} \) is that the \( K \) function is an accumulative statistic and \( O_{12} \) is noncumulative. Therefore, the spatial pattern at small scales does not directly influence the \( O \) statistic at larger scales, whereas the cumulative \( K \) function confounds effects at larger distances with effects at shorter distances (Getis and Franklin 1987). The present study requires using a noncumulative statistic because we search for critical scales in the spatial patterns of blowouts along utilization gradients by livestock.

We used a grid-based estimator of the O-ring statistic such as that used by Wiegand et al. (1999) and Condit et al. (2000) and explained in detail by Wiegand and Moloney (2004). The spatial resolution was in accordance to our raster data, 30 × 30 m, and we used a ring width of \( w = 150 \) m to avoid jagged plots of \( O_{12} \) (which result if not enough blowouts fall within the rings). In our study the type 1 points were water points and the type 2 points were blowouts. We analyzed the area covered by the lightly grazed paddocks and that covered by the heavily grazed paddocks separately.

In a first step we calculated for each paddock \( i \), the number of blowouts in rings with fixed width \( w = 150 \) m and variable...
radius $r$ centered at the corresponding water point ($= p_{i2}[r]$) and the area of dune crest within these rings ($= a_i[r]$; Fig. 1). Note that blowouts and dune crests in the part of the rings outside the respective paddock do not count here (Fig. 1) because livestock could not move from one paddock to the other (e.g., livestock located at paddock H1 did not contribute to formation of the utilization gradient of paddock H2). In a second step we averaged over all four paddocks with the same treatment to obtain the estimator of the $O$-ring statistic:

$$O_{12}(r) = \frac{(1/4) \sum_{i=1}^{4} p_{i2}(r)}{(1/4) \sum_{i=1}^{4} a_i(r)}$$

To explore local effects within an individual paddock $i$ we also calculated the individual $O$-ring statistic:

$$O_{12}^i(r) = \frac{p_{i2}(r)}{a_i(r)}$$

To find out if and at which spatial scale blowouts were significantly aggregated around water holes we contrasted the empirical $O$-ring statistic to simulation envelopes constructed from Monte Carlo simulations of a null model that randomized the locations of the blowouts over the area of dune crests, thereby removing potential spatial correlation between water points and blowouts. Because water points preceded blowouts, the interaction can occur in only one direction; thus we kept the locations of the water holes unchanged.

The approximate 95% simulation envelopes were the 25th lowest and highest value of $O_{12}(r)$, taken from the 999 simulations of the null model. At a given distance $r$, a value of $O_{12}(r)$ below or above the simulation envelop is interpreted as significant attraction or repulsion, respectively. However, the simulation envelopes cannot be interpreted as confidence intervals in a strict sense. Because of simultaneous inference (i.e., we tested all spatial scales $r$ simultaneously), type I error may occur if the value of $O_{12}(r)$ is close to a simulation envelope (i.e., the null model may be rejected even if it is true; Loosmore and Ford 2006). However, although this is a special concern with the accumulative $K$ function, it is less a concern with noncumulative statistics (Loosmore and Ford 2006). Therefore, we interpret our results, especially small departures from the null model, with caution.

To find out if the density of dune crests around water points influences the soil erosion processes, we estimated the intensity of dune crests using a circular moving-window estimate with bandwidth of $R = 60$ m (Bailey and Gatrell 1995). The intensity of dune crests at a given location was given by the number of points that are located within a neighborhood with radius $R = 60$ m, divided by the area of this neighborhood. If a neighborhood was partly outside the irregularly shaped study region, the number of points in the incomplete neighborhoods was divided by the proportion of the area that lay within the study region. We then classified the intensity of dune crests (measured in dune crests $\cdot 30 \times 30$ m cells$^{-1}$) roughly into low (intensity $< 0.1$), medium (intensity $< 0.6$), and high (intensity $> 0.6$). Water points were classified according to their location into low, medium, and high dune crest–density water points. The densities of blowouts in the 300-m neighborhood of these three types of water points (given by the $O$-ring statistic) were compared by analysis of variance, for both lightly and heavily grazed paddocks.

Figure 1. Study area and example for the calculation of the $O$-ring statistics in paddock H3. The irregularly shaped study area is represented by a light grey color, dune crests are represented as dark grey, blowouts are shown as black squares, and water points are shown as white disks. The four lightly grazed paddocks are labeled as L1, L2, L3, and L4, and the four heavily grazed paddocks are labeled as H1, H2, H3, and H4. For the calculation of the spatial statistics, rings with width $w = 150$ m and variable radius, $r$, are centered in the water hole of paddock $i$ and the number of blowouts (white squares; $= p_{i2}[r]$) and the number of cells with dune crests (dark grey; $= a_i[r]$) in the segment of the circles that fall inside the paddock are counted.
RESULTS

A total of 171 blowouts were identified for the study area. On the mosaic of aerial photos (year 1969) 157 blowouts were digitized, but 14 new blowouts were identified on the Landsat image (year 2002). The area occupied by smaller blowouts on the mosaic and on the image corresponded with the minimum mapping unit for each sensor ($0.36 \times 10^{-2}$ ha and $9 \times 10^{-2}$ ha, respectively). Therefore, blowouts generated after the year 1969 with areas smaller to $9 \times 10^{-2}$ ha could not be mapped.

The lightly grazed paddocks contained 63 blowouts and the heavily grazed paddocks, 108 (Fig. 1). The overall density of blowouts within dune crest areas was 0.52 and 0.28 blowouts $\cdot$ ha$^{-1}$ at the lightly and heavily grazed paddocks, respectively (gray horizontal lines in Fig. 2). Point-pattern analysis using the O-ring statistic reveals that blowouts at the lightly grazed paddocks showed weak small-scale aggregation (up to 30 m) around water points and at distances of about 300 m from the water point (Fig. 2A). In contrast, in heavily grazed paddocks the aggregation of blowouts was strong at distances between 90 and 210 m from water points and significant at distances of 60–390 m. Blowout density peaked at a distance of 180 m, showing a density threefold higher than expected at random (i.e., the horizontal line in Fig. 2). In lightly grazed paddocks the aggregation was much weaker. Interestingly, density of blowouts did not peak at the heavily grazed paddock close to the water point, but some 200 m away. This result shows that the activity of livestock may have the strongest impact not immediately close to the water hole but a "walking distance" away.

To explore our data in more detail and to evaluate possible site effects, we also analyzed the data from each paddock separately using the individual O-ring statistics (equation 2). Clearly, because of the low sample sizes, stochastic effects were large. Nevertheless, we found in all four heavily grazed paddocks significant aggregation of blowouts at some distances from water points shorter than 400 m (Appendix, Fig. A1). In contrast, only one lightly grazed paddock showed aggregation of blowouts at distances shorter than 100 m (Appendix, Fig. A2, panel D). The density of dune crests in relation to water point locations is shown in Figure 3. In two paddocks (H1 and L4), water points were located in areas with high density of dune crests, whereas in paddocks H4 and L1 water points were in areas with very low density of crests. In the remaining four paddocks, the water points were in areas with medium intensity of dune crests. To find out if the density of blowouts differed among water holes with low, intermediate, and high densities of dune crest we conducted two analyses of variance, separately, for lightly and heavily grazed paddocks. As measure of the density of blowouts we used the first 10 distance classes of the corresponding individual O-ring statistics. In lightly grazed paddocks blowout density increased systematically with density of dune crests, but we found a significant difference only between high density of dune crests and the low and intermediate class (Fig. 4A). In heavily grazed paddocks blowout density also increased systematically with density of dune crests, but we found a significant difference between low density of dune crests and the intermediate and high class (Fig. 4B). Thus, under light grazing, blowout density increased significantly only if water points were located in areas with a high density of dune crests, whereas blowout density significantly increased in heavily grazed paddocks if water points were located in areas with an intermediate density of dune crests.

DISCUSSION

By applying spatial analysis to data derived from remotely sensed imagery, detailed spatial information can be inferred when trying to understand the dynamic of soil erosion processes within desert rangelands. Furthermore, the indicator that we used, presence or absence of blowouts, is suitable for dealing with soil erosion processes, because blowouts reflect a...
large temporal scale and also are quickly and easily detected in remote sensing images.

We found aggregated patterns of blowouts around water points in the vegetated dune field at Peninsula Valdés, with significant aggregation under heavy grazing pressure. Under these conditions, sheep may severely affect soil stability through forage intake, trampling, and/or concentration in resting points (Evans 1998). Resting, which may at first glance represents a nondestructive, low-impact activity, can affect vegetation and soil stability if resting places are located in fragile sites (Evans 1997). Indeed, some of the blowouts found in the study area seem to have originated in resting places (bare areas with a high density of sheep feces). This may weaken our findings; however, the overall signature of aggregation of blowouts in the neighborhood of water points was strong enough to be significant. Other animals commonly mentioned as initiators of sandy soil surface erosion are rabbits ( Ranwell 1960; Jungerius and Van der Meulen 1988), cows ( Thom et al. 1992), bears ( Martini 1981), and horses ( Lucas 1988).

The aggregated pattern of blowouts around water points was stronger and occurred at larger distances under heavy sheep grazing compared with light sheep grazing. Strong aggregation occurred for heavily grazed paddocks between 90 and 210 m. The absence of an aggregated pattern of blowouts close to the water point (less than 100 m) could be because of the nature of pisospheres. This innermost zone is usually trampled to dust because of heavy grazing disturbance ( Trash and Derry 1999). Similar aggregation of blowouts around water points was found in other desert rangelands grazed by sheep ( Nechaeva 1979). This author found an aggregated pattern of blowouts around water points at a larger distance (maximum of 3 500 m) in Central Asia sand rangelands. Grazing based in transhumance in Russian rangelands, which implies radius of sheep migration of about 5 000 to 6 000 m ( Mainguet 1991), could explain the longer distance of the grazing effect.

In the established paddocks of Patagonia the animals graze at distances up to 3 000 m from water points ( Bisigato and Bertiller 1997).

Our results also suggest that responses of soil erosion processes to grazing impact contrast among sites with different sensitivity to accelerated erosion. As the availability of vulnerable sites to erosion increased (high density of dune crests around water points), grazing impacts on rangeland became more intense irrespective of the historical stocking rate. However, at a medium density of dune crests, the effects of grazing on blowout formation were only significant in the heavily grazed paddocks (Fig. 4). Increases of susceptibility to soil erosion in zones with a high density of dune crests may be explained by the topographic conditions (heights and depressions) where aerodynamic flows are laterally and horizontally compressed and accelerated, especially on the top of the dunes ( Carter 1988). The approach taken here could even be
extended to map the susceptibility of soil erosion in a qualitative way, based on the three variables of grazing regime, distance from water point, and dune crest density.

MANAGEMENT IMPLICATIONS

We found that grazing effects on soil erosion processes in the vegetated dune field at Península Valdés is conditioned by the spatial heterogeneity of sensitive sites to dune reactivation. Our findings suggest that the location of water points in sites less susceptible to suffering accelerated soil erosion should significantly improve with strategies based on reductions in stock. Sensitive areas can be fenced off to exclude livestock. Choices of fencing types could include sand fencing, which has an added component of facilitating the filling in of gaps in dune crests (Nordstrom 2000). Sand fences reduce wind speed near the ground, thereby causing moving sand to be deposited into a mound on the downwind side of the fence and reducing deposition farther downwind (Gares 1990). Our conclusion has significant management implications in desert rangelands, where the proposed management solutions have usually focused on the need to reduce the total number of livestock grazing.

ACKNOWLEDGMENTS

We wish to acknowledge the valuable suggestions made by Matthew Bowker and three anonymous reviewers. Also, Diego Giberto provided valuable suggestions that greatly improved the manuscript. As well, we gratefully acknowledge Fernando Coronato and Graciela Metternicht, who revised the English. Comisión Nacional de Actividades Espaciales from Argentina supplied the satellite images.

LITERATURE CITED


Figure A1. Blowout spatial pattern around water points in heavily grazed paddocks. • = ring statistic (blowouts · ha⁻¹); black lines indicate the upper and lower limits of the 95% simulation envelope of the null model. Points above the envelope indicate aggregation and points below the envelope indicate regulate pattern. The horizontal lines indicate the overall density of blowouts.
Figure A2. Blowout spatial pattern around water points in lightly grazed paddocks. \( O_{12}(r) \) statistic (blowouts \( \cdot \) ha\(^{-1} \)); black lines indicate the upper and lower limits of the 95% simulation envelope of the null model. Points above the envelope indicate aggregation and points below the envelope indicate regulate pattern. The horizontal lines indicate the overall density of blowouts.